Presence of Noroviruses and Other Enteric Viruses in Sewage and Surface Waters in The Netherlands

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Since virus concentrations in drinking waters are generally below the detection limit, the infectious risk from drinking water consumption requires assessment from the virus concentrations in source waters and removal efficiency of treatment processes. In this study, we estimated from reverse transcription-PCR on 10-fold serially diluted RNA that noroviruses, the most prevalent waterborne gastroenteritis agents, were present at 4 (0.2 to 38) to 4,900 (303 to 4.6×10^4) PCR-detectable units (PDU) per liter of river water (ranges are given in parentheses). These virus concentrations are still high compared with 896 to 7,499 PDU/liter of treated sewage and 5,111 to 850,000 PDU/liter in raw sewage. Sequencing analyses designated human norovirus GGII.4 Lordsdale as the most prevalent strain in the sampling period 1998 to 1999 in both sewage and surface waters. Other GGII strains were also very abundant, indicating that the majority of the virus contamination was derived from urban sewage, although very divergent strains and one animal strain were also detected in the surface and sewage waters. Rotaviruses were also detected in two large rivers (the Maas and the Waal) at 57 to 5,386 PDU/liter. The high virus concentrations determined by PCR may in part be explained by the detection of virus RNA instead of infectious particles. Indeed, reoviruses and enteroviruses that can be cultured were present at much lower levels, of 0.3 to 1 and 2 to 10 PFU/liter, respectively. Assuming 1% of the noroviruses and rotaviruses to be infectious, a much higher disease burden than for other viruses can be expected, not only because of the higher levels but also because of these viruses' higher infectivity and attack rates.

Noroviruses (NoV; previously known as Norwalk-like caliciviruses or small round-structured viruses) (35) belong to the most infectious group of causative agents of epidemic gastroenteritis (11, 44). Genetically, noroviruses can be divided into five genogroups (GGI, GGII, GGIII, GGIV, and GGV), which consist of different genotypes. A prototype virus represents each genotype: GGI includes Norwalk virus (GGI-1) and Southampton virus (GGI-2), GGII includes Hawaii virus (GGII-1) and others (2, 57), GGIII includes Jena virus, GGIV includes Alphatron, and GGV is found in mice (25, 53).

Numerous outbreaks of norovirus-associated gastroenteritis originating from contaminated drinking or recreational water have been ascribed to various causes (4, 6, 21, 29, 34, 37, 40). Contamination of drinking water by sewage through pump failure or blockage of a sewage system has been previously described (6, 34). Also, inadequate or failing treatment processes have led to the insufficient removal of viruses from source waters (4, 29, 37). Waterborne outbreaks may arise from direct exposure by ingestion of contaminated tap water or water-containing products, e.g., ice cubes, custard, and salads. Outbreaks associated with the consumption of contaminated drinking water may go unnoticed because of compliance with guidelines, including bacterial indicators and state-of-the-art treatment processes (17). In waterborne outbreaks, a very high proportion of the population can be affected, leading to several

to hundreds of cases of gastroenteritis, followed by secondary spread and resulting in significant economic impact.

Noroviruses have been detected in raw urban sewage (32, 33). Primary and secondary sewage treatment processes do not efficiently reduce the virus concentration, in contrast with tertiary processes (13, 24, 46, 52). Therefore, depending on the applied processes, treated sewage discharged onto surface waters may significantly enhance the virus concentrations in the environment. Other sources for viral contamination of surface waters may occur during heavy rainfall and include discharges of raw sewage or washoff of animal manure. The latter may give rise to public health problems in the case of exposure to zoonotic pathogens, as was suggested occurs with caliciviruses and hepatitis E viruses (3, 18). Drinking water may be produced from groundwater or surface waters. Surface waters are more heavily contaminated with pathogenic microorganisms, although viruses have also been detected in groundwater (8, 14, 16, 46). The efficiency of advanced drinking water treatment processes may differ significantly with regard to removal or inactivation of microorganisms from source waters. Bacteria are more readily inactivated by chlorination than are protozoa (49). Viruses are effectively removed or inactivated by slow sand filtration and soil passage, but they are more resistant to UV and coagulation combined with sedimentation (9, 43, 50). However, a very low concentration of virus may suffice to cause infection, e.g., from 10 to $<10^4$ norovirus PCR-detectable units (PDU), leading to gastrointestinal disease in two-thirds of the individuals infected (30). Quantitative risk assessment based on source water quality and removal efficiency of treatment processes is required with regard to pathogenic viruses, since direct detection of pathogenic viruses in drinking water is not possible. This study focused on noroviruses, which are the most

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FIG. 1. Sampling locations in The Netherlands. Grey fields indicate urbanization; thick lines indicate rivers. Numbers indicate sampling locations. 1, Sewage treatment plant at Apeldoorn, raw sewage; 2, sewage treatment plant at Apeldoorn, treated sewage; 3, river Waal; 4, river Maas.

common waterborne gastroenteritis agents in The Netherlands. We estimated the numbers of particles containing norovirus and rotavirus RNA in two large rivers by molecular methods. We compared these estimates with the numbers of infectious enterovirus and reovirus particles and bacteriophages in these surface waters. Noroviruses in raw and treated sewage waters were quantified and typed by sequencing analysis and compared to strains detected in the surface waters to identify possible sources of contamination. Finally, norovirus sequences originating from water samples were compared to consensus strains circulating in the population.

MATERIALS AND METHODS

Viruses. The positive control samples used in the cell culture method were virus isolates, reovirus type 3, and coxsackievirus type B4, originally obtained from the Voorns Canal and grown by buffalo green monkey kidney (BGM) cell culture. The positive control used in the detection of F-specific bacteriophages was MS2 (ATCC 15597-B1), and the positive control in the detection of somatic coliphages was φ X174 (ATCC 13706-B1). Norovirus-positive stool specimens obtained from patients with gastroenteritis were used as a positive control (GGII.1; Hu/NV/6649/1994) for RNA extraction and reverse transcription-PCR (RT-PCR). Rotavirus WA was used as a positive control for RNA extraction and

RT-PCR (Diagnostic Laboratory for Infectious Diseases, National Institute of Public Health and Environment).

Water samples. Ten liters of raw sewage and 200 to 400 liters of treated sewage were collected at a pumping-engine station in the city of Apeldoorn, The Netherlands (Fig. 1). The treatment process consists of primary settlement of the sewage, biological treatment (with activated sludge), and phosphorus removal. Large volumes of water (200 to 600 liters) were collected from two large Dutch rivers, the Maas and the Waal. Sampling at all locations was repeated at approximately 4-week intervals in the period from November 1998 to April 1999.

Concentration of the water samples. Water was first concentrated by a conventional filter adsorption-elution method. The average recovery of the membrane filtration step used in this study was 87 to 94% as determined by the use of F-specific bacteriophages (these phages have characteristics comparable to the NoVs with respect to size and isoelectric point) and 80% for coxsackie B4 virus. Briefly, for the concentration procedure, magnesium chloride was added to the water sample to a final concentration of 0.05 M to enable formation of a virus-magnesium chloride complex. By reducing the pH to 3.8, these positively charged complexes adsorb to a negatively charged cartridge filter (Nominal; 1.2 μm-pore-sized filter) (Millipore, Etten-Leur, The Netherlands). Viruses were eluted from the filter with a 3% beef extract (Difco, Amsterdam, The Netherlands) solution, to which Tris (Biosolve, Valkenswaard, The Netherlands) was added to a final concentration of 0.05 M to set the pH to 9.0. The typical retentate volume of 10-liter raw sewage water was approximately 650 ml, and the volume of large volumes of treated sewage and surface water (200 to 600 liters) was approximately 1,800 ml. The precipitate was dissolved and neutralized with a concentrated acetic acid buffer (pH 5.0) to a final pH of approximately 7.2. The resulting eluate in part was subjected further to the two-phase separation method and in part was analyzed by cell culture for enteric viruses and bacteriophages.

Concentration and purification of virus from the eluate by two-phase separation. The two-phase separation method is based on the selective distribution of viruses into two incompatible phases (42), with slight modifications. Briefly, 650 ml of eluate, 1% (wt/vol) Dextran T40 (Pharmacia, Roosendaal, The Netherlands), 10% (wt/vol) polyethylene glycol 6000 (Merck, Amsterdam, The Netherlands), 0.2 M NaCl (Merck), and 10 mM phosphate buffer (pH 7.2) were mixed for 1 h at 4°C on a horizontal shaker. The suspension was then transferred to a separation funnel and left overnight at 4°C. After separation, the bottom phase and the interphase were harvested. Further purification was done by spin column gel chromatography with Sephadex G200 (ICN, Zoetermeer, The Netherlands) and by ultrafiltration with a Centricon 100 microconcentrator (100,000-molecular-weight cutoff; Amicon, Dronten, The Netherlands). The average retentate volumes were 1 to 5 ml.

RNA extraction. RNA was extracted from the retentate obtained after two-phase separation by binding to silica beads in the presence of a high-molarity solution (5). This method was modified to allow larger volumes of retentate to be processed (32). Briefly, 2.7 M solid guanidinium isothiocyanate (Fluka, Zwijndrecht, The Netherlands), 25 mM EDTA (pH 8.0; ICN), 55 mM Tris-HCl (pH 6.4; Biosolve), and 20 μl of silica suspension were added to the retentate, resulting in a final volume of 1.7 to 8.7 ml. The sample was mixed on a rotary shaker, and then centrifuged briefly to pellet the silica particles. The pellet was washed twice with a guanidinium isothiocyanate-containing washing buffer (L2) (5), twice with ethanol 70% (vol/vol), and once with acetone. After removal of acetone by evaporation, the RNA was eluted in distilled water with RNAguard (200 U/ml; Pharmacia) and dithiothreitol (3 mM; Sigma, Zwijndrecht, The Netherlands) and was used directly in RT-PCR.

RT-PCR. For norovirus detection, we used an RT-PCR assay that had been developed for the specific detection of norovirus in stool specimens from humans (56). The assay detects over 85% of a panel of different antigenic types of norovirus (27). Norovirus-specific primers were used, which anneal to a part of the norovirus genome within the viral RNA polymerase gene (open reading frame 1). The RT-PCR was done as described previously (54, 56). For rotaviruses, we used an RT-PCR assay developed for generic molecular detection, amplifying the VP6 gene fragment as described previously (55). RNA was diluted at 10-fold intervals to semiquantitatively determine the virus concentrations in the samples.

Gel electrophoresis, Southern blotting, and hybridization. The amplification products were analyzed by electrophoresis with 2% agarose gels and visualized under UV illumination after being stained with SYBR Gold nucleic acid gel stain (Molecular Probes, Leiden, The Netherlands). The PCR products in the agarose gel were transferred to a positively charged nylon membrane (Amersham, 's Hertogenbosch, The Netherlands) by a vacuum blotting system (Millipore) with 0.5 M NaOH (Merck) and 0.6 M NaCl (Merck) for 30 min. The confirmation of the specificity of the detected RT-PCR products of both viruses was done as described previously (54, 55).

Norovirus cloning, sequencing, and phylogenetic analysis. Since water may contain multiple virus strains (32), each norovirus-positive RT-PCR product was cloned and selected for sequence analysis. Norovirus hybridization-positive RT-PCR products were excised from a 2% agarose gel and purified with a Qiaquick gel extraction kit (QIAGEN, Hilden, Germany). The purified RT-PCR products were cloned into the pGEM-T Easy Vector System II (Invitrogen, Leek, The Netherlands); after transformation, at least five positive colonies were selected. The pGEM-T Easy Vector system was checked for correct insertion size by direct PCR amplifications with M13 forward and M13 reverse primers supplied by the manufacturer. PCR products hybridizing with the generic norovirus probes (1, 56) were purified with a PCR purification kit (QIAGEN) and sequenced with the BigDye Terminator Cycle Sequencing Ready Reaction kit (Perkin-Elmer, Applied Biosystems, Foster City, Calif.). Nucleotide sequences were edited by SeqMan II (DNA-star) and aligned by Bionumerics, version 2.0 (Applied Maths, Kortrijk, Belgium) by the unweighted-pair group method using average linkages after multiple sequence alignment of a 145-nucleotide segment of the polymerase gene.

Cell culture. Virus infectivity was determined by a monolayer plaque assay (7). Briefly, BGM cells were grown to confluent monolayers in 75-cm² plastic flasks. The culture medium was removed, after which the eluate and antibiotic mixture were added to the flasks. The cultures were incubated at room temperature for 120 min to allow virus adsorption to the cells. The cells were overlaid with Medium 199 with Earl's salts (Life Technologies, Breda, The Netherlands) with 10% fetal bovine serum (Life Technologies), 0.9% Bacto agar (Difco, Amsterdam, The Netherlands), 0.2% bicarbonate, 100 IU of penicillin, and 100 μg of

streptomycin (Life Technologies)/ml. After 9 days of incubation at 37°C, the cells were stained with 0.03% neutral red in 0.9% agar. After 24 h, the plaques were enumerated and the virus concentration in the original water sample was calculated from the test volume and the virus count.

Detection of bacteriophage. The detection of F-specific bacteriophages and somatic coliphages was done according the ISO 10705-1 (1995) and ISO 10705-2 (2000) standards (www.iso.org), respectively. The host bacterial strains WG49 and WG5 were exposed to the eluates to culture the F-specific and somatic bacteriophages, respectively (20). In the presence of bacteriophages, plaques could be enumerated after overnight culture. The virus concentration in the original water sample was calculated from the test volume and the virus count.

Statistical methods. The numbers of virus particles present in water were estimated by culture or by RT-PCR on 10-fold serially diluted RNA (end point dilution). Virus concentrations in the undiluted samples were estimated as most probable numbers by the use of the number of PFU or the presence or absence of virus genomes in the 10-fold RNA dilutions, under the assumption that negative samples do not contain virus or viral RNA. Application of the Poisson distribution was justified by the assumption that the infectious virus particles or viral RNA was dispersed randomly in the sample. The maximum likelihood method was used to estimate the number of virus particles in the undiluted sample (38). A negative binomial model gave the best fit for the distribution of virus particles in the original and diluted samples. The 95% confidence interval was estimated for each virus concentration.

RESULTS

Detection of viruses in surface water samples. Virus concentrations in surface water samples were determined for samples from two large rivers in the centre of The Netherlands (Fig. 1). These rivers, the Maas and Waal Rivers, are influenced by (un)treated sewage originating from Belgium and France, and Germany, respectively. Each river water sample was positive for the presence of F-specific and somatic phages, noroviruses, rotaviruses, reoviruses, and enteroviruses (Table 1). In the Maas, the average concentrations of F-specific and somatic phages were 7×10^3 and 2×10^4 PFU/liter, respectively. The average reoviruses and enterovirus concentrations were much lower, varying from 1 to 6 PFU/liter. On average, noroviruses and rotaviruses were both detected at 2×10^2 PDU/liter in the Maas. In the Waal, the average concentrations of F-specific and somatic phages were 1×10^3 and 3×10^3 10⁴ PFU/liter, respectively. As observed for the Maas, the concentrations of reoviruses and enteroviruses were much lower (0.3 to 10 PFU/liter), whereas noroviruses and rotavirus concentrations were higher (on average, 2×10^3 PDU/liter).

Detection of norovirus variants in surface water. Strains belonging to seven different norovirus genotype strains (GGI.2) Southampton, GGI.4 QueensArms, GGII.3 Mexico, GGII.3R Rotterdam, GGII.4 Lordsdale, GGII.7 Leeds, and an unclassified GGII strain) were found in the surface water samples out of a total of 38 clones. Figure 2 shows these different norovirus strains detected in surface water and the relation of these strains to the consensus of the prototypes found in the population (28). As can be seen from Fig. 2, most norovirus strains from the water samples were closely related to the respective prototypes, differing from 0% (GGII.4 Lordsdale) to 12% (GGII.7 Leeds). However, one strain was found to differ by over 15% in two clones isolated from either the Maas or Waal River sampled on the same day (28 December 1998). The aberrant strain (designated Maas/Waal) was compared to consensus strains of variants found in fecal samples from patients in the population. Aberrant or other variants were circulating in the population years later and varied over 15% with the strains found in the surface water samples (Fig. 3a). To deter-

In Institute I. Detection of norovirus, rotavirus, reovirus, enterovirus, somatic phages, and F-specific phages in river water samples, Maas and Waal Rivers

			Concu	Concn (95% confidence interval)	rval)		
Date	Location	Somatic phages (PFU/liter)	F-specific phages (PFU/liter)	Enteroviruses (PFU/liter)	Reoviruses (PFU/liter)	Noroviruses (PDU/liter)	Rotaviruses (PDU/liter)
10 December 1998 10 December 1998 28 December 1998 28 December 1998 28 January 1999 28 January 1999 11 March 1999 11 March 1999	Maas Waal Maas Waal Maas Waal Maas	4.0 × 10 ⁴ (3.1 × 10 ⁴ to 4.3 × 10 ⁴) 6,963 (6,231–7,751) 1.8 × 10 ⁴ (1.2 × 10 ⁴ to 2.7 × 10 ⁴) 7.7 × 10 ⁴ (7.0 × 10 ⁴ to 8.5 × 10 ⁴) 2.8 × 10 ⁴ (2.5 × 10 ⁴ to 3.1 × 10 ⁴) 2.1 × 10 ⁴ (1.8 × 10 ⁴ to 2.3 × 10 ⁴) 1.6 × 10 ⁴ (1.3 × 10 ⁴ to 1.9 × 10 ⁴) 5,305 (4,523–6,173)	$1.5 \times 10^4 (1.4 \times 10^4 \text{ to } 1.6 \times 10^4)$ 319 (279-362) 1.572 (1.28-1.960) 609 (541-682) 4.460 (4.070-4.874) 2.386 (2.141-2.648) 6.054 (5.530-6.610) 1.282 (1.159-1.413)	1 (0.7–2) 0.6 (0.3–1) 1 (0.9–2) 0.3 (0.1–0.6) 1 (0.7–2) 0.3 (0.1–0.6) 2 (1–3) 0.8 (0.2–2)	6 (5-8) 6 (5-7) 5 (4-7) 5 (4-6) 4 (3-6) 10 (9-12) 3 (2-5) 2 (0.8-4)	85 (5-756) 776 (49-6,925) 4 (0.2-38) 4,900 (303 to 4.6 × 10 ⁴) 397 (19-3,554) 29 (1-257) 370 (18-3,304) 540 (26-4,819)	57 (3–504) 5,386 (333 to 5.1 × 10 ⁴) 285 (13–2,556) 315 (20–2,823) 264 (13–2,369) 193 (12–1,730) 248 (12–2,220) 360 (17–3,213)

mine the recombination or mutation rate of these strains, a comparison was made between the sequences found in the two surface waters and the sequences of these consensus strains. Figure 3b shows the differences in nucleotides between these strains. In each surface water sample, strains were detected which belonged to at least one norovirus genotype strain, but sometimes two or three strains. In all four Maas samples, strains were found grouped with four different norovirus genotypes (QueensArms, Mexico, Lordsdale, and the Maas/Waal strain) and in the river Waal samples, five different norovirus genotypes (Southampton, Rotterdam, Lordsdale, Leeds, and the Maas/Waal strain) were found. In samples from both rivers, strains were present from the Lordsdale and the aberrant GGII strain genotypes. The additional strains were found in several samples and several sequenced clones. On two of four sampling dates in either river, strains genotyped as Southampton, QueenArms, and Mexico were found. The strains belonging to the three additional genotypes (Rotterdam, Leeds, and the Maas/Waal strain) were found each in only one out of the four sampling dates and were found scattered over the sampling period.

Detection of viruses in sewage samples. Raw and treated sewage samples were collected every 3 to 4 weeks during the winter of 1998 to 1999. F-specific and somatic phages, enteroviruses, reoviruses, rotaviruses, and noroviruses were detected in all sewage samples (Table 2).

In raw sewage, the concentrations of F-specific and somatic phages were higher than that found in river water (average, 10^6 PFU/liter). On average, reoviruses and enteroviruses were present at 10^2 PFU/liter, whereas noroviruses and rotaviruses, respectively, were detected on average at 2×10^5 and 2×10^4 PDU/liter of raw sewage. The average virus concentrations in treated sewage were lower than in raw sewage, except for rotaviruses (Table 2).

Virus removal at the sewage treatment plant was on average 1.6 and 1.1 \log_{10} units for F-specific and somatic phages, respectively. For enteroviruses and reoviruses, removal was similar at 1.4 and 1.3 \log_{10} units on average, respectively. For noroviruses, the average removal was 1.8 \log_{10} units, whereas for rotaviruses the removal was much lower, approximately 0.2 \log_{10} units (Table 2).

Typing of norovirus variants in sewage waters. To distinguish between different norovirus strains present in water, we determined the genetic variability of these noroviruses by sequencing the RT-PCR products after cloning. The sequences of 145 nucleotides were aligned and subjected to phylogenetic analysis, with 86 clones sequenced in all. The sequences amplified from water samples were compared to consensus strains found in samples from the human population.

Strains belonging to eight different norovirus genotype strains (GGI.2 Southampton, GGII.1 Hawaii, GII.2 Melksham, GGII.3 Mexico, GGII.4 Lordsdale, GGII.7 Leeds, an unclassified GGII strain [Apeldoorn RS01-11-99], and bovine NoV) were found in 10 sewage samples (Fig. 4). As can be seen from Fig. 4, most norovirus strains from the sewage samples are closely related to the respective prototypes differing from 0% (GGII.4 Lordsdale and GGII.3 Mexico) to 12% (GGII.7 Leeds), except one strain differing by over 15% (Apeldoorn RS01-11-99). In each sewage sample, strains were detected which belonged to at least one norovirus genotype strain, but

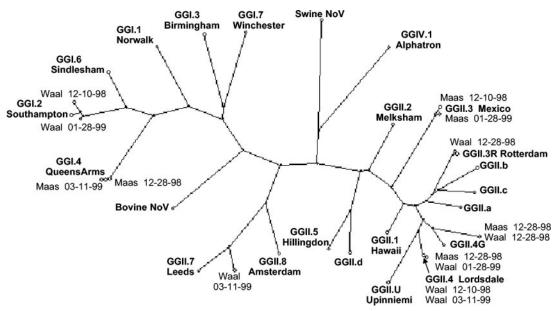


FIG. 2. Phylogenetic tree illustrating the genetic relationship of norovirus isolates from river water samples from the Maas and Waal Rivers and the consensus sequences of norovirus prototypes (GG) found in the population (28). The genetic relationship is based on a 145-nt fragment of the polymerase gene.

sometimes two or three different strains. In the raw sewage samples, strains were found belonging to six different norovirus genotypes strains in all (GGI.2 Southampton, GGII.1 Hawaii, GII.2 Melksham, GGII.3 Mexico, GGII.4 Lordsdale, and the Apeldoorn RS01-11-99 strain). In the treated sewage samples, strains were found belonging to five different norovirus genotype strains (GGI.2 Southampton, GGII.3 Mexico, GGII.4 Lordsdale, GGII.7 Leeds, and bovine NoV). Strains belonging to the genotypes Southampton, Mexico, and Lordsdale were predominantly present in both raw and treated sewage. The strains belonging to the other genotypes were retrieved from 1 or 2 of the 86 sequenced clones. On all sampling dates, strains belonging to the genotypes Southampton and Lordsdale were present in either raw or treated sewage or both. On three of five sampling dates, strains belonging to the genotype Mexico were found in raw sewage or treated sewage. The strains belonging to the five additional genotypes (Hawaii, Melksham, Leeds, the Apeldoorn RS01-11-99 strain, and bovine NoV) were found scattered over the sampling period.

DISCUSSION

Surface waters may be directly or indirectly contaminated with human pathogenic viruses by (un)treated sewage or washoff of animal manure. We found noroviruses, rotaviruses, and enteroviruses in each of the samples taken from two large rivers in The Netherlands. In the summer of the same year, Schvoerer et al. (45) did not detect rotaviruses, adenoviruses, or hepatitis A viruses in fresh waters in France by PCR, but in 1 of 10 samples, noroviruses were detected when 5 ml was analyzed; in 3 of 10 samples, enteroviruses were detected. In another French study, viruses were detected by PCR and culture methods in concentrates of 20 liters of river water (23). From the 68 samples tested, only 2 were positive for infectious enterovirus but 60 were positive by enterovirus RT-PCR. Also

other viruses were detected with PCR, one sample was positive for hepatitis A virus, one sample was positive for norovirus genogroup II, and three samples were positive for astrovirus. From 11 of 22 800-ml river water samples in Japan, 100 reovirus strains were identified by electropherotyping (36). In comparison, we detected reoviruses by cell culture in each of the eight river water samples tested. By RT-PCR, three of eight 1-liter water samples from three rivers in Germany were positive for the presence of enteroviruses in 1993 (15). Two years later, each of six 1-liter river water samples tested positive for the presence of GGI noroviruses, rotaviruses, and enteroviruses; in two of six samples, GGII noroviruses were detected, but no hepatitis A viruses were detected (16). In a recent Finnish study, 1-liter surface water samples were concentrated, of which 13 of 139 samples tested positive for the presence of noroviruses and 47 of 139 samples tested positive for the presence of F-RNA phages (22). These few papers document presence-absence data; however, to be able to estimate the infectious risk by consumption of drinking water, quantitative data are necessary (19). We determined the virus concentrations varying from 0.3 to 10 PFU of enteroviruses and reoviruses/liter of river water and varying from 319 to 7.7 \times 10⁴ PFU of F-specific and somatic phages/liter. Since noroviruses and rotaviruses are hard or impossible to culture (12), these viruses were detected by molecular methods resulting in much higher concentrations, varying from 4 to 5,386 PDU/liter of river water. No sensitive cell lines are available; alternatively, the presence of the opposite strand of these singlestrand viruses could be determined to establish if replication took place. This has been described for adenoviruses but yet has to be developed for noroviruses (26). In absence of such methods, we estimated the virus concentrations by RT-PCR on 10-fold serial dilutions of the extracted RNA as a semiquantitative approach. A negative binomial model was used to esti-

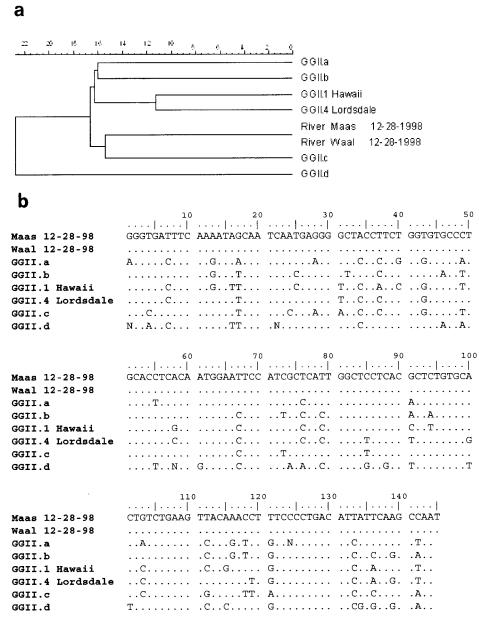


FIG. 3. (a) Phylogenetic analysis of the polymerase gene (145-bp region) showing the aberrant Maas/Waal strain (River Maas 12-28-98; River Waal 12-28-98) detected in the river water, Maas and Waal Rivers, and different GGII consensus prototype strains. (b) Alignment of the polymerase gene (145-bp region) showing the aberrant Maas/Waal strain (Maas 12-28-98; Waal 12-28-98) detected in the river water, Maas and Waal Rivers, and different GGII consensus prototype strains.

mate the virus concentration in the original sample because this model gives the best fit for the distribution of virus particles in the samples (51).

The Maas is a river fed by rainwater. In total, 8 million people live in the Maas catchment area, with over 5 million people depending on it for drinking water, compared with 30 million people who drink water directly or indirectly produced from the river Rhine. The Rhine is a river fed by both rain and meltwater. Our results generally showed higher concentrations of pathogenic viruses in the Waal, a tributary of the Rhine, than in the Maas. The Rhine catchment is five times as large as the Maas, collecting more microbiological contamination from

wastewater sources. Since part of the population in the catchments is not yet connected to sewage treatment and part of the sewage is discharged without treatment with (for instance) heavy rainfall, this would indicate that high virus loads may reach the surface waters. Indeed, we found loads varying from 111 to 2,143 PFU of enterovirus and reovirus/liter of raw sewage, 6.8×10^5 to 6.8×10^6 PFU of F-specific and somatic phages/liter, and 339 to 8.5×10^5 PDU of rotaviruses and noroviruses /liter. In treated sewage, we determined virus concentrations varying from 5 to 92 enterovirus and reovirus PFU/ liter, 6.2×10^3 to 1.5×10^5 PFU of F-specific and somatic phages/liter, and 598 to 2.4×10^4 PDU of rotaviruses and

				Concn (95% confidence interval)	fidence interval)		
Date	Sewage	Somatic phages (PFU/liter)	F-specific phages (PFU/liter)	Enteroviruses (PFU/liter)	Reoviruses (PFU/liter)	Noroviruses (PDU/liter)	Rotaviruses (PDU/liter)
.6 November 1998 .6 November 1998	Raw Treated	$3.5 \times 10^6 (2.9 \times 10^6 \text{ to } 4.2 \times 10^6)$ $3.6 \times 10^5 (2.9 \times 10^5 \text{ to } 4.4 \times 10^5)$	$\begin{array}{llllllllllllllllllllllllllllllllllll$	191 (140–253) 39 (31–49)	1,253 (1,114–1,403) 48 (39–59)	7,035 (444 to 6 $.3 \times 10^4$) 896 (57–7,999)	$4,690 (296 \text{ to } 4.2 \times 10^4)$ 598 (38–5,333)
Calculated log 10 removal			1.6	0.7	1.4	0.9	0.9
21 December 1998 21 December 1998	Raw Treated	$6.8 \times 10^{6} (6.3 \times 10^{6} \text{ to } 7.3 \times 10^{6})$	$6.8 \times 10^6 (6.9 \times 10^3 \text{ to } 7.7 \times 10^5)$ 6.7 × 10 ³ (5.3 × 10 ³ to 7.7 × 10 ³)	288 (226–361) 7 (4–10)	2,143 (1,964–2,332)	1.2×10^{5} (7,359 to 1.1×10^{6}) 1 367 (88 to 1.2 × 10 ⁴)	763 (36–8,804) 908 (57–8 106)
Calculated log 10 removal			2.0	1.6	1.4	1.9	-0.1
1 January 1999	Raw	$6.4 \times 10^6 (5.0 \times 10^6 \text{ to } 8.1 \times 10^4)$	$1.2 \times 10^6 (1.1 \times 10^6 \text{to} 1.3 \times 10^6)$	296 (186–424)	555 (414–725)	$8.5 \times 10^5 (5.3 \times 10^4 \text{ to } 8.0 \times 10^6)$ $5.5 \times 10^4 (2,592 \text{ to } 4.9 \times 10^5)$	$5.5 \times 10^4 (2,592 \text{ to } 4.9 \times 10^5)$
1 January 1999	Treated	$1.0 \times 10^6 (9.3 \times 10^5 \text{to} 1.1 \times 10^6)$	$6.7 \times 10^4 (5.9 \times 10^4 \text{ to } 7.6 \times 10^4)$	10 (7–13)	76 (68–85)	7,499 (463 to 7.1×10^4)	$4,819 (228 \text{ to } 4.3 \times 10^4)$
Calculated log 10 removal		0.8	1.3	1.5	0.9	2.1	1.1
February 1999	Raw		$7.5 \times 10^5 (6.7 \times 10^5 \text{ to } 8.4 \times 10^6)$	140 (99–190)	416 (343–499)	4.6×10^4 (2,896 to 4.1×10^4)	$3.1 \times 10^4 (1,931 \text{ to } 2.7 \times 10^5)$
February 1999	Treated	$1.5 \times 10^5 (1.3 \times 10^6 \text{to} 1.7 \times 10^5)$	$1.3 \times 10^4 (1.2 \times 10^4 \text{to} 1.4 \times 10^4)$	5 (3-7)	23 (19–28)	$4,283 (265 \text{ to } 4.0 \times 1 0^4)$	$2.9 \times 10^4 (1,763 \text{ to } 2.7 \times 10^5)$
Calculated log 10 removal		1.4	1.8	1.5	1.3	1.0	0.03
25 February 1999	Raw	$5.2 \times 10^6 (4.5 \times 10^6 \text{ to } 6.0 \times 10^6)$	$3.5 \times 10^6 (3.2 \times 10^6 \text{ to } 4.0 \times 10^6)$	833 (587–1,141) 111 (40–239)	111 (40-239)	$5{,}111 (322 \text{ to } 4.6 \times 10^4)$	339 (16–3,027)
25 February 1999	Treated	Treated $1.9 \times 10^5 (1.8 \times 10^8 \text{ to } 2.1 \times 10^5)$	$5.4 \times 10^4 (4.8 \times 10^4 \text{ to } 6.2 \times 10^4)$ 12 (9–18)	12 (9–18)	8 (5–12)	$3,582 (221 \text{ to } 3.4 \times 10^4)$	$2.4 \times 10^4 (1,475 \text{ to } 2.2 \times 10^5)$
Calculated log 10 removal		1.4	1.8	1.8	1.1	0.2	-1.8

TABLE 2 Detection of norovirus, rotavirus, reovirus, enterovirus, somatic phages, and F-specific phages in sewage samples (both raw and treated sewage) and the virus removal at sewage treatment plant the

noroviruses/liter. The contribution of the sewage treatment processes to virus removal varied between 0.7 and 2.1 \log_{10} units. This indicates that discharge of both raw and treated sewage will considerably contribute to the virus burden of the rivers.

Noroviruses are genetically very diverse because of accumulating mutations during replication (57). Sequencing analysis of norovirus PCR products showed the presence of 10 different genotypes in sewage and surface waters. Half of these genotypes were detected in both sewage and surface waters, indicating that these virus genotypes were very prevalent in The Netherlands. In addition, the diverging genotypes that were detected either in sewage at the same plant or in rivers several tens of kilometers away pointed at regional differences between genotypes circulating in smaller groups of the population or sporadic cases.

Recently, the circulation of viruses in the population and environment was shown for enteroviruses (47). In 1998 to 1999, mainly GGII strains were found in fecal samples taken from humans in The Netherlands (27). In both sewage and surface waters, GGII strains were also predominantly detected. More specifically, GGII.4 Lordsdale was the main genotype in both the Dutch human population and waters, followed by GGII.3 Mexico, GGII.7 Leeds, and GGII.1 Hawaii. The predominant GGI strain in sewage, surface waters, and the population was GGI.2 Southampton. Remarkably, a bovine NoV strain was detected in urban sewage, known as the Jena variant (31). However, in both sewage and surface waters aberrant strains were found which were not represented in clinical samples. One diverging variant (Apeldoorn RS01-11-99 strain) was detected in raw sewage in January 1999, which differed significantly from any variant identified to that point. The highest identity of 85% in the 145-nucleotide segment of the polymerase gene was found for GGIIU Upinniemi. Another divergent strain (Maas/Waal) detected in both the rivers Maas and Waal showed the highest homology (86%) with GGIIc Den Haag. At least four different GGII.4 Lordsdale strains emerged in 1998 to 1999, causing gastroenteritis in the population.

Noroviruses are generally more prevalent in winter (39). Here, we screened sewage and surface waters sampled in the cold season in which noroviruses were present. However, the screening period was not very extensive, ranging only from November to April. We have also detected noroviruses in water samples taken in warmer periods of the year at other locations, which is relevant when surface waters are used for recreational purposes (10, 48). We have also previously shown that viruses occur at peak concentrations, suggesting that the risk of infections may also differ greatly between exposures (52, 58). Moreover, consumption of drinking water may also pose a health risk if it coincides with failed drinking water treatment. Exposure to high levels in surface waters by either recreation, shellfish culture, or drinking water production leads to public health hazards. Although a low dose of only 10 PDU may lead to infection (30), not all of these virus units making up the high-virus concentrations found in surface waters will lead to infection, because of host factors. But even if only 1% of noroviruses detected in surface waters are infectious, this would lead to a much higher disease burden in case of noro-

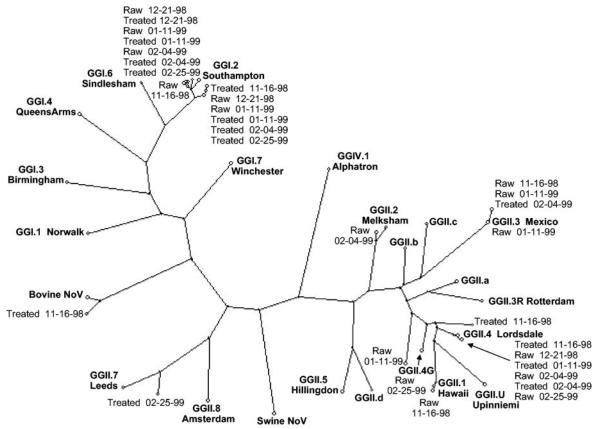


FIG. 4. Phylogenetic tree illustrating the genetic relationship of norovirus isolates from sewage samples (both raw and treated sewage) and the consensus sequences of norovirus prototypes (GG) found in the population (28). The genetic relationship is based on a 145-nt fragment of the polymerase gene.

viruses (attack rate, 65% gastroenteritis cases independent of blood group) (41) than for enteroviruses or reoviruses.

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